



Baseline

Metal concentrations in the beach sediments of Bahía Solano and Nuquí along the Pacific coast of Chocó, Colombia: A baseline study



Harry Gutiérrez-Mosquera^{a,d}, V.C. Shruti^b, M.P. Jonathan^{b,*}, Priyadarsi D. Roy^c, D.M. Rivera-Rivera^b

^a Facultad de Ingeniería, Universidad Tecnológica del Chocó, Carrera 22 No.18B-10, Quibdó, Colombia

^b Centro Interdisciplinario de Investigaciones y Estudios sobre Medio Ambiente y Desarrollo (CIEMAD), Instituto Politécnico Nacional (IPN), Calle 30 de Junio de 1520, Barrio la Laguna Ticomán, Del. Gustavo A. Madero, C.P. 07340 Ciudad de México, Mexico

^c Instituto de Geología, Universidad Nacional Autónoma de México (UNAM), Ciudad Universitaria C.P. 04510, Del. Coyoacan, Ciudad de México, Mexico

^d Facultad de Ingeniería, Universidad de Medellín, Carrera 87 No. 30-65, Medellín, Colombia

ARTICLE INFO

Keywords:
Beach sediment
Acid leachable metal
Geochemical index
Potential ecological risk
Chocó
Colombia

ABSTRACT

Thirty sediment samples from four different beaches along Bahía Solano and Nuquí (Department of Chocó) of eastern Colombia, with tourism and gold mining activities, were analysed to estimate the concentrations of fourteen different acid leachable metals. Metal distribution patterns showed elevated concentrations of Co, Cr, Cu, Pb and Zn compared with the upper continental crust values. Calculation of geochemical indices confirmed that the enrichment is due to periodic gold mining activities (severe to extremely severe enrichment of Cu, Zn, V, Co, Cr and Pb) along with natural (geological) contributions (minor and moderate enrichment of Ca, Mg, Fe, Ti, Mn and Li). Potential ecological risk index revealed that Pb posed the highest risk. Our results together with a global comparison suggest that the observed metal enrichments are mainly caused by mining and to a lesser extent by tourism in this region, thus instigating continuous monitoring of metal concentrations in this region.

Tourism is considered as one of the biggest global industries with beaches hosting free plays and people as potential revenue partners. Beach tourism results in large economic benefits, thereby increasing the demand for new construction of high rise buildings, resorts, cottages and recreation parks at beaches (Propín-Frejomil and Sánchez Crispín, 2007; Faggi and Dadon, 2011; Pérez-Maqueo et al., 2017). The impact of tourism invasion has often led to the disposal of various contaminants, thereby causing serious concerns to coastal ecosystem (Vikas and Dwarkish, 2015; Wang et al., 2016). Metals are one among the most toxic contaminants that have received considerable attention on global scale owing to their close relationship with human health (Chopra and Pathak, 2015; Ye et al., 2015; Ramachandra et al., 2018). They enter the coastal waters through various natural (weathering and erosion) and anthropogenic (mining, waste disposal, agricultural and industrial activities) sources, wherein they do not remain in soluble form for a long time and have a tendency to deposit in sediments (Xu et al., 2016; El Nemr et al., 2016). Therefore, sediments, being the potential storehouse of metals, are the best tools to assess the contamination status of a given aquatic ecosystem.

In the recent years, our research group has been exploring this storehouse with the vision of generating baseline data for

understanding metal enrichments in various beaches located worldwide. We succeeded in generating a database for popular tourist destinations such as Mexico (Acapulco, Huatulco and Santa Rosalía), India (Chennai), South Africa (Richards Bay, South Durban, Sodwana Bay and St Lucia) and Malaysia (Miri), which have different geological characteristics. Acid leachable technique was effectively adopted as the fundamental method to determine metal concentrations for evaluating the potential environmental and ecotoxicological impacts in beach regions.

In this context, our vision now focuses on Colombia, where beach tourism symbolises one of the most important revenue income. During the period of 2009–2011, nearly 974,721 international (mostly from the U.S.A., Canada and the European Union) and 3,411,523 domestic tourists visited the coastal areas of Colombia (PROEXPORT - Ministerio de Comercio, Industria y Turismo, 2011). The capacity of rapid growth of this sector in the coming decades will bring economic strength to this region. Among diverse coastal regions, Chocó in Colombia is characterised with a unique ecosystem and unexploited natural resources. The richness of Chocó's landscape and its enormous biodiversity make this region quite overwhelming. It is the only region of Colombia that is lapped by the waves of two oceans: the Pacific and the Caribbean. In

* Corresponding author.

E-mail address: mpjonathan7@yahoo.com (M.P. Jonathan).

particular, the coast of Bahía Solano and Nuquí forms the economic and touristic corridor of Chocó. It offers plethora of activities such as scuba diving and sport fishing. This tourist capital of the Pacific coast of Colombia attracts a large number of visitors to explore the marine life, beaches and deep jungles with all the natural biodiversity that is different from other local destinations. Its additional economy relies on artisanal fisheries, agriculture and timber extraction (Matallana, 1999; Avila et al., 2008). Mining has preceded this region since the Spanish colonisation (Leyva, 1993). Numerous studies have focused on massive and uncontrolled gold mining activities in Chocó (Vallejo Toro et al., 2016; Gutiérrez-Mosquera et al., 2017; Salazar-Camacho et al., 2017; Palacios-Torres et al., 2018) and in adjacent regions (Olivero et al., 2002; Marrugo-Negrete et al., 2008; Cordy et al., 2011; Alvarez et al., 2012a, 2012b). Given this scenario, we ask whether the intense economic development associated with tourism, urban sprawl and mining are driving any geochemical changes to the beaches of Bahía Solano and Nuquí. On this background, the framework of this study involves the evaluation of metal concentrations in beach sediments along Bahía Solano and Nuquí to comprehend the natural and external (human induced) impacts on the marine ecosystem.

The investigation was carried out on a ~200-km-long coastal strip of Bahía Solano and Nuquí in Chocó department of eastern Colombia ($6^{\circ}19'40''N/77^{\circ}22'41.20''W$ and $5^{\circ}36'28.20''N/77^{\circ}26'28.40''W$). A total of thirty sediment samples were collected from different beaches during May 2016. Based on sample collection, the coastal stretch was categorised into four different regions, wherein 1 and 2 belong to the municipality of Bahía Solano, and 3 and 4 belong to Nuquí municipality: 1) Mutis beaches (S. Nos. 1–13), 2) El Valle beaches (S. Nos. 14–21), 3) Nuquí beaches (S. Nos. 22–26) and 4) Coqui beaches (S. Nos. 27–30) (Fig. 1). Region 1 hosts a port for cargo boats and passenger transportation services and stores trading provisions (materials and timber). Region 2 has a “fishing village” at the mouth of the Río Valle, where fertile breeding ground results in abundant fishing activities. Region 3 is characterised with mosaic ecosystems such as hot springs, rivers and beaches. It also contains a protective coastal marine environment (i.e. Utría National Natural Park) with humpback whales and sea turtle nests. Region 4 hosts the best-preserved mangroves of Colombia and offers excellent ecotourism. The geology of all the four regions comprises tholeiitic basalt, dolerite, basic tuff and volcanic breccia of the late cretaceous, limestone, sandstone and calcareous mudstone of the Oligocene-Miocene, alluvial deposits of the Quaternary and gravel, and sand and mud rich in organic materials of the Holocene (Gómez Tapia and Almanza Meléndez, 2015).

Sediment samples were collected from the intertidal zone using a

plastic spatula and placed in clean polythene bags. The collected samples were then transported to the laboratory and oven dried below 40 °C. Coning and quartering methods were applied to obtain a homogenous part of the sample, and the sample was powdered using an agate mortar for further analyses. Approximately 1 g of the sample was digested according to the modified EPA 3051A method (2007) and as described by Navarrete-López et al. (2012) using 2.5 ml of HNO₃, 0.8 ml of HCl and 1 ml of H₂O₂ acids (all analytical grade) in a closed vessel, which was made of a resistant material Poly(tetrafluoroethylene) (PFA) at 119 ± 1.5 °C for 40 min. The final solution was made up to 10 ml after filtration and analysed for concentrations of fourteen metals (Al, Fe, Mg, Ca, Ti, Co, Cr, Cu, Mn, Pb, Zn, V, Sr and Li) using an inductively coupled plasma optical emission spectrometer (ICP-OES 8300 PerkinElmer). Standard reference materials namely SRM No. 691029, Loam soil B and Marine Reference Standard (SRM 2702 – Inorganics in marine sediment) were run along with the samples. The recoveries were as follows: Al 97.89% ($\pm 1.62 \mu\text{g g}^{-1}$), Fe 98.82% ($\pm 1.72 \mu\text{g g}^{-1}$), Mg 110.88% ($\pm 1.30 \mu\text{g g}^{-1}$), Ca 98.01% ($\pm 1.65 \mu\text{g g}^{-1}$), Ti 97.39% ($\pm 2.50 \mu\text{g g}^{-1}$), Co 101.39% ($\pm 3.10 \mu\text{g g}^{-1}$), Cr 102.35% ($\pm 2.10 \mu\text{g g}^{-1}$), Cu 101.66% ($\pm 3.20 \mu\text{g g}^{-1}$), Mn 92.70% ($\pm 0.08 \mu\text{g g}^{-1}$), Pb 111.07% ($\pm 2.81 \mu\text{g g}^{-1}$), Zn 95.93% ($\pm 1.10 \mu\text{g g}^{-1}$), V 106.01% ($\pm 1.55 \mu\text{g g}^{-1}$), Sr 105.46% ($\pm 2.10 \mu\text{g g}^{-1}$) and Li 110.42% ($\pm 1.05 \mu\text{g g}^{-1}$). The detection limits for the analysed metals were Al 0.01%, Fe 0.01%, Mg 0.01%, Ca 0.01%, Ti 0.01%, Co 1 ppm, Cr 2 ppm, Cu 1 ppm, Mn 1 ppm, Pb 2 ppm, Zn 1 ppm, V 1 ppm, Sr 1 ppm and Li 0.0001–1%.

The distribution pattern of analysed metals is presented in Fig. 2 a–n. The descending order of metal concentrations (all values in $\mu\text{g g}^{-1}$) in all the four regions is as follows: 1) Mutis beach: Fe (34,111) > Ca (9871) > Al (8846) > Mg (6494) > Ti (1699) > Pb (391) > Mn (372) > Cr (239) > V (176) > Zn (125) > Co (108) > Sr (49) > Cu (44) > Li (15); 2) El Valle beach: Fe (34,581) > Al (16,839) > Mg (15,070) > Ca (6152) > Ti (801) > Mn (556) > Pb (399) > Cr (279) > V (143) > Zn (127) > Co (111) > Cu (85) > Sr (51) > Li (30); 3) Nuquí beach: Fe (25,067) > Mg (15,788) > Al (10,684) > Ca (9388) > Ti (1099) > Pb (430) > Mn (416) > Cr (302) > Zn (124) > Co (109) > V (104) > Cu (36) > Sr (28) > Li (17) and 4) Coqui beach: Fe (20,875) > Mg (9661) > Ca (8068) > Al (7735) > Ti (963) > Pb (467) > Cr (309) > Mn (284) > Zn (130) > Co (110) > V (83) > Cu (45) > Sr (30) > Li (16). The above-mentioned array displayed no significant variations in metal concentrations in all the four regions. For a better understanding of the enrichment pattern, metal concentrations were compared with the upper continental crust



Fig. 1. Locations from which sediments were collected in four different beaches along Bahía Solano and Nuquí, Department of Chocó, Colombia.

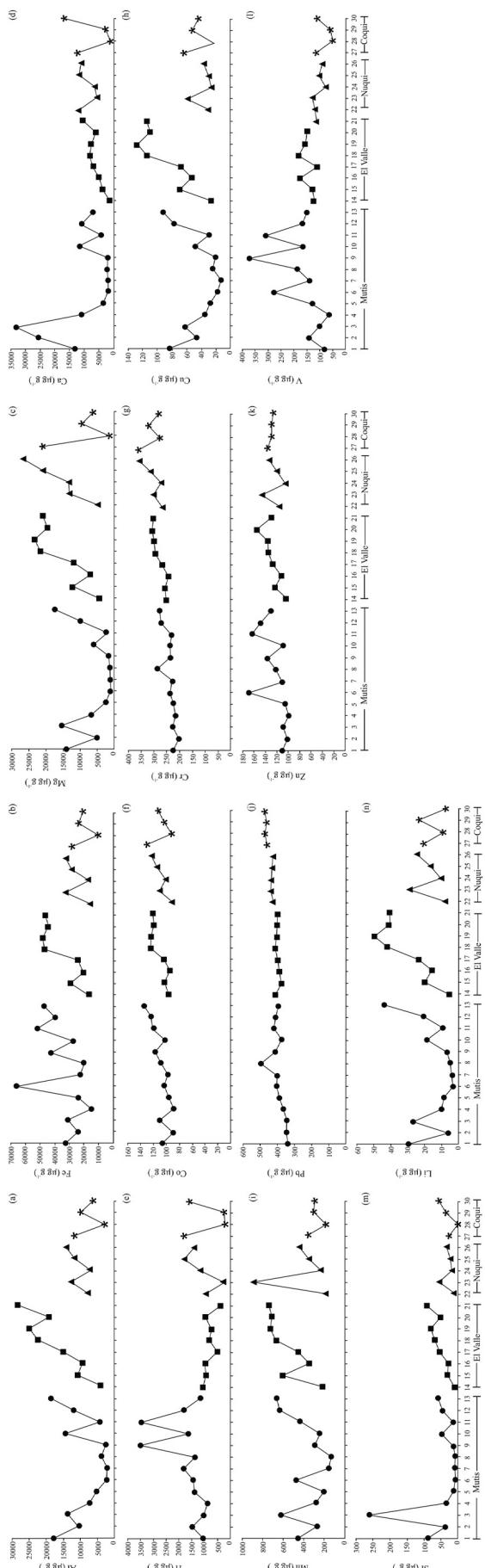


Fig. 2. a-n. Concentrations of different metals in beach sediments collected from four different regions along Bahía Solano and Nuquí, Department of Chocó, Colombia.

(UCC; Wedepohl, 1995) background values. Extreme enrichment levels (all values in $\mu\text{g g}^{-1}$) for Co (108; 111; 109; 110), Cr (239; 279; 302; 309), Cu (44; 85; 36; 45), Pb (391; 399; 430; 467), Zn (125; 127; 124; 130) and V (176; 143; 104; 83) in all the four regions were witnessed when compared with the UCC background values (Table 1).

Spatial variability in metal concentrations (all values in $\mu\text{g g}^{-1}$) revealed that the majority of metals such as Al (16,839; 10,684), Fe (34,581; 25,067), Mg (15,070; 15,788), Co (111; 109), Cr (279; 302), Cu (85; 36), Mn (556; 416), Pb (399; 430), Zn (127; 124) and Li (30; 17) are elevated in El Valle and Nuquí beaches (Fig. 2 a-n). Elevated levels of aforementioned metals are geological, and they are attributed to the alluvial deposits of the Quaternary and organic material-rich Holocene sediments (Gómez Tapias and Almanza Meléndez, 2015). Higher values of elements such as Ca (9871) and Sr (49) of the Mutis beach are derived from the carbonate-bearing sedimentary rocks (limestone and calcareous mudstone) of the Oligocene-Miocene (Gómez Tapias and Almanza Meléndez, 2015). Additional geological component of this region is the presence of volcanic deposits (tholeiitic basalt, basic tuff and volcanic breccia) (Gómez Tapias and Almanza Meléndez, 2015), which consequently contributed to higher concentrations of V (176) and Ti (1699). Compared to other regions, elevated concentrations of Cu (85) were observed in El Valle beaches. This can be explained by contributions from Cu in antifouling paints and impregnation of nets used in fish farming (Vallejo Toro et al., 2016). Gold production in this region accounts for 23.6% of Colombia's total production and represents 15% of Chocó's GDP (Goñi et al., 2014). A considerable number of tailings are inevitable during the retrieval of precious metals (gold, silver and platinum). These mine tailings are generally characterised with higher concentrations of Co, Cr, Cu, Pb, Zn and V (Ferreira da Silva et al., 2004; Bryan et al., 2006; Ahmadi et al., 2015; Fashola et al., 2016; Falagán et al., 2017). Apart from the natural geological inputs, the higher levels (avg. values in $\mu\text{g g}^{-1}$) of Co (109), Cr (269), Cu (54), Pb (410), Zn (126) and V (143) along the entire coast are attributed to the mining activities related to platinum, gold and silver production in the Department of Chocó. San Juan and Dagua are important rivers that drain into the Pacific coast of Chocó further south to the study area and have been significantly affected by gold mining activities (Maria Fernanda and Collazos-Santos, 2014). The distribution of metals in sediments along the Pacific coast of Chocó is strongly influenced by the longshore sand drift. It has a dominant northeast component during rainy season (April–May) when southerly winds are dominant in some sectors and set up short, high-frequency waves (Correa and Morton, 2011; Rangel-Buitrago et al., 2017). In addition, the Humboldt Current system flowing in the direction of the equator transports the metal-contaminated sediments bought in by the rivers (San Juan & Dagua) from the south to the north of the Chocó.

The inter-relationships among different metals were identified by correlation analysis. Statistica 8 software was used wherein the entire dataset was Varimax normalised and the Pearson correlation matrix was generated with $p < 0.5$, 0.01 and 0.001 (Table 2). The data analysis presented in this study represents only significantly correlated values, which have some direct relationship in the geochemical process. The significant correlation of Al and Fe vs Co ($r^2 = 0.54$; 0.67), Cu ($r^2 = 0.93$; 0.47), Mn ($r^2 = 0.73$; 0.67), Zn ($r^2 = -$; 0.76) and Li ($r^2 = 0.93$; 0.48) suggested their detrital origin associated with aluminosilicates from the alluvial deposits and organic-rich sediments of the Quaternary. Ca and Sr presented ($r^2 = 0.73$) a strong positive correlation, and the assemblage of Ca-Sr combination is believed to behave similarly with each other due to their similar ionic radii ($r_{\text{Sr}} = 0.113 \text{ nm}$ vs $r_{\text{Ca}} = 0.099 \text{ nm}$) and charge. Sr is typically found in calcium-bearing minerals, and their strong inter-relationship reflects the presence of carbonate rocks such as limestone, sandstone and calcareous mudstone in the study region (Gómez Tapias and Almanza Meléndez, 2015). In addition, the strong inter-relationship among Co vs Cu ($r^2 = 0.57$), Zn ($r^2 = 0.61$), Cr ($r^2 = 0.61$) vs Pb ($r^2 = 0.63$) suggested their common origin. The correlation among Co, Cu, Zn, Cr and Pb indicates the

Table 1
Comparison of concentrations of the studied metals in the sediments of Bahía Solano and Nuquí with those in other global coastal environments.

Locations	Extraction type	Metal concentrations						References
		Al	Fe	Mg	Ca	Ti	Co	
Worldwide								
Portugal coast	HCl + HNO ₃ + HF	—	12,000–88,000	—	—	—	22–89	18–133
Rosetta coast, Egypt	HCl + HNO ₃ + HF	—	109,560	—	—	—	69,78	0.18
Salaam coast, Tanzania	HCl + HNO ₃	—	461–5352	—	—	—	0.21–2.75	1–9.6
Huatulco beaches, Mexico	HCl + HNO ₃	—	7492	—	—	—	4.63	89.26
Venezuela coast	—	—	—	—	—	—	—	10.27
Arabian Gulf, Saudi Arabia	—	—	3447	—	—	—	6.35	27.10
Chennai, India	HCl + HNO ₃	442	—	—	—	—	—	14.10
Lutong beach, Malaysia	HCl + HNO ₃	1888	—	—	—	—	12.64	85.33
Quintero Bay, Chile	—	—	—	—	—	—	—	45.7
Korea coast	HCl + HNO ₃ + HF	67,000	43,700–95,800	6000–15,500	2700–37,500	—	9.6	58.3
Gulf of Urabá, Colombia	HCl + HNO ₃ + HF	61,800–134,200	—	—	—	—	—	109.61–212.45
Stratoni_Ierissos Gulf, Greece	XRF	—	—	—	—	—	—	—
Sulcis-Iglesiente, Sardinia Italy	HCl + HNO ₃	—	—	—	—	—	—	5.59
Sodwana Bay, South Africa	HCl + HNO ₃	—	7784	1008	—	—	—	426
Present study								
Bahía Solano Beaches, Colombia	HCl + HNO ₃	11,136	30,964	10,752	8558	1261	109.08	269.17
Sediment quality guidelines								
TEC	—	—	—	—	—	—	—	43.4
PEC	—	—	—	—	—	—	—	111
Ecotoxicological values								
LEL	—	—	20,000	—	—	—	—	26
SEL	—	—	40,000	—	—	—	—	110
ERL	—	—	—	—	—	—	—	81
ERM	—	—	—	—	—	—	—	370
Locations	Metal concentrations						References	
	Cu	Mn	Pb	Zn	V	Sr	Li	
Worldwide								
Portugal coast	117–884	200–4632	35.7–437	38.3–349.3	—	—	—	
Rosetta coast, Egypt	24.57	55.3	384.68	183.23	374.78	—	114.06	
Salaam coast, Tanzania	0.3–2.1	17–219	0.8–2.2	2.6–9.3	1.1–13.7	—	—	
Huatulco beaches, Mexico	3.35	69–24	4.48	7.86	—	—	—	
Venezuela coast	3.79	—	11.64	113.38	—	—	—	
Arabian Gulf, Saudi Arabia	5.78	75.17	58.68	17.57	—	—	—	
Chennai, India	4.05	46.8	19.77	9.89	—	—	—	
Lutong beach, Malaysia	28.81	12.83	12.88	17.93	—	—	—	
Quintero Bay, Chile	464	602	25.2	111	—	—	—	
Korea coast	36.5	35	122	—	—	—	—	
Gulf of Urabá, Colombia	25.08–102.94	361.92–905.47	0.17–6.93	75.15–161.53	—	—	—	
Stratoni_Ierissos Gulf, Greece	143	10,150	1146	1863	—	—	—	
Sulcis-Iglesiente, Sardinia Italy	2.87	204,33	203	3272	—	—	—	
Sodwana Bay, South Africa	4.53	72,16	1,27	2,81	—	—	—	
Present study								
Bahía Solano Beaches, Colombia	53.93	416.82	409.67	125.80	142.55	43.59	19.37	

(continued on next page)

Table 1 (continued)

Locations	Metal concentrations						References
	Cu	Mn	Pb	Zn	V	Sr	Li
Sediment quality guidelines							
TEC	31.6	—	35.8	121	—	—	MacDonald et al., 2000
PEC	149	—	128	459	—	—	
Ecotoxicological values							
LEL	16	460	31	120	—	—	USEPA, 2001
SEL	110	1100	250	820	—	—	
ERL	34	—	46.7	150	—	—	Long et al., 1995
ERM	270	—	218	410	—	—	

All values expressed in ($\mu\text{g g}^{-1}$). UCC = upper continental crust (Wedepohl, 1995); TEC: threshold effect concentration; PEC: probable effect concentration; LEL: lowest effect level; SEL: severe effect level; ERL: effect range low; ERM: effects range medium.

presence of chalcophile elements and the influence of external anthropogenic sources in the study area. The long-term gold mining activities contribute to the source of Co, Cr, Cu, Pb and Zn metals in the beach sediments of Bahía Solano and Nuquí.

To have a panoramic view, the metal concentration data from other tourist beaches worldwide along with that from Bahía Solano and Nuquí beach sediments are presented in Table 1. Vallejo Toro et al. (2016), Pappa et al. (2016) and Romano et al. (2017) reported significant impacts of gold mining activities on metal concentrations in the coastal sediments of the Gulf of Urabá (Colombia), Stratoni_Ierissos Gulf (Greece) and Sulcis-Iglesiente (Sardinia Italy), respectively. In the present study ($\mu\text{g g}^{-1}$), Cr (269), Cu (54), Pb (410) and Zn (126) concentrations were within the range or even lower compared to those in other regions. From the results of the present study and a previous study on the Gulf of Urabá (Colombia), we report that both regions namely the Atlantic coast and Pacific coast of Colombia are greatly affected by extensive gold mining activities. Similarly, when compared with metal concentrations in other tourist beaches such as Huatulco (Mexico), Chennai (India) and Lutong (Malaysia), those obtained in the present study were approximately 10- to 15-fold higher. El-Sorogy et al. (2016), Rumisha et al. (2012), García et al. (2008), Youssef et al. (2015) and Ra et al. (2013) demonstrated that anthropogenic activities such as urbanisation, industrialisation and sewage disposals greatly influence the metal concentrations in the sediments of Rosetta coast (Egypt), Salaam coast (Tanzania), Venezuela coast, Arabian Gulf (Saudi Arabia) and Korea coast, respectively. Compared with these regions, Bahía Solano and Nuquí sediments had higher concentrations (avg.) of Co (109.08), Cr (269.17), Cu (53.93), Pb (409.67) and Zn (125.80). Globally, Bahía Solano and Nuquí beach sediments presented higher concentrations of metals than other tourist and human-influenced beaches. However, the results of the present study were found similar to those obtained in areas influenced by gold mining activities, thus confirming the substantial impact of gold mining activities on the metal concentrations of Bahía Solano and Nuquí sediments.

We compared the obtained metal concentrations with sediment quality guidelines (SQGs) (MacDonald et al., 2000) and ecotoxicological values (USEPA, 2001; Long et al., 1995) to evaluate the quality of sediments. SQGs and ecotoxicological values were developed to evaluate the degree to which the sediment-associated metal concentrations might adversely affect the aquatic biota and help in the interpretation of the sediment quality. They include the threshold effect concentration (TEC), probable effect concentration (PEC), lowest effect level (LEL), severe effect level (SEL), effect range low (ERL) and effect range medium (ERM). The values below TEC, LEL and ERL indicate no harmful effects while values above PEC, SEL and ERM suggest the frequent occurrence of adverse biological effects. In this study, Cr ($269 \mu\text{g g}^{-1}$) and Pb ($410 \mu\text{g g}^{-1}$) concentrations in Bahía Solano and Nuquí sediments pose severe threats to the aquatic biota because they exceed the PEC, SEL and ERM values (Table 1).

Further, the degree of metal contamination in sediments was assessed using geochemical indices such as the enrichment factor (EF) geoaccumulation index (I_{geo}) and potential ecological risk index (PERI). We used UCC (Wedepohl, 1995) values for calculating the geochemical index. The enrichment factor distinguishes between crustal and non-crustal sources, thereby estimating the anthropogenic sources of metals in sediments (Buat-Ménard and Chesselet, 1979; Ibrahim et al., 2014; Ashraf et al., 2016). It encompasses the normalisation of metal concentrations with uncontaminated background values: $EF = (\text{Me}/\text{Al})_{\text{sample}}/(\text{Me}/\text{Al})_{\text{baseline}}$, where $(\text{Me}/\text{Al})_{\text{sample}}$ and $(\text{Me}/\text{Al})_{\text{baseline}}$ represent the ratio of metal to Al concentration in the studied samples and in the background sample, respectively. Al is highly abundant in earth's crust, and we used it as a normaliser in this study. Moreover, its close association with aluminosilicates makes it one of the dominant metal-bearing phases in sediments (Windom et al., 1989; Schropp et al., 1990; Alexander et al., 1993; Vink et al., 1999; Gao and Chen, 2012). Different classes and their corresponding EF values calculated for the

Table 2

Correlation matrix analyses of sediments collected from different beaches along Bahía Solano and Nuquí, Department of Chocó, Colombia.

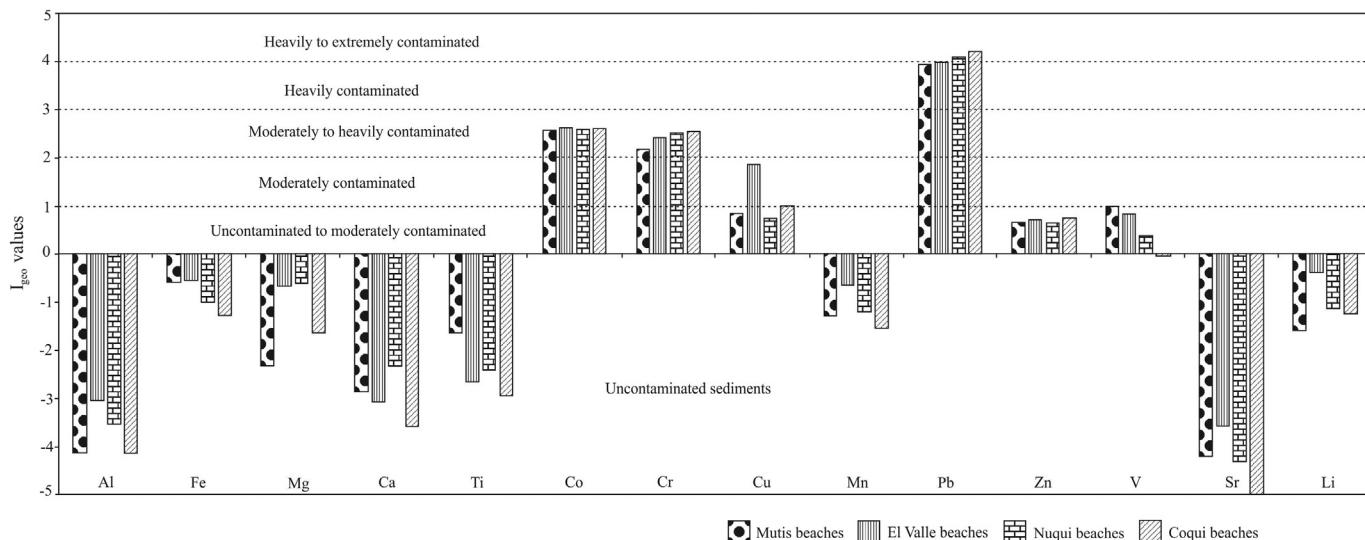
	Al	Fe	Mg	Ca	Ti	Co	Cr	Cu	Mn	Pb	Zn	V	Sr	Li
Al	1.00													
Fe	0.40*	1.00												
Mg	0.82*†‡	–	1.00											
Ca	–	–	–	1.00										
Ti	–0.41*	–	–	–	1.00									
Co	0.54*†	0.67*†‡	0.66*†‡	–	–	1.00								
Cr	0.39*	–	0.69*†‡	–	–	0.61*†‡	1.00							
Cu	0.93*†‡	0.47*†	0.71*†‡	–	–0.38*	0.57*†	–	1.00						
Mn	0.73*†‡	0.67*†‡	0.65*†‡	–	–	0.63*†‡	–	0.79*†‡	1.00					
Pb	–	–	–	–0.44*	–	–	0.63*†‡	–	–	1.00				
Zn	–	0.76*†‡	–	–	–	0.61*†‡	0.38*	–	0.53*†	–	1.00			
V	–	0.62*†‡	–	–	0.73*†‡	–	–	–	–	–	0.52*†	1.00		
Sr	0.52*†	–	0.42*	0.73*†‡	–	–	–	0.49*†	0.51*†	–0.45*	–	–	1.00	
Li	0.93*†‡	0.48*†	0.81*†‡	–	–0.41*	0.65*†‡	0.46*†	0.94*†‡	0.81*†‡	–	–	–	0.51*†	1.00

p < 0.05*, 0.01†, 0.001‡.

Table 3

Calculated enrichment factor (EF) values for sediments collected from the beaches along Bahía Solano and Nuquí, Department of Chocó, Colombia.

Beaches	Enrichment factor values (EF)						
	No enrichment (< 1.5)	Minor enrichment (< 3)	Moderate enrichment (3–5)	Moderately severe enrichment (5–10)	Severe enrichment (10–25)	Very severe enrichment (25–50)	Extremely severe enrichment (> 50)
Mutis	Sr	Ca	Mg	Ti, Mn, Li	Fe	Cu	Co, Cr, Pb, Zn, V
El Valle	Ca, Sr	Ti		Fe, Mg, Mn, Li	Zn, V	Cu	Co, Cr, Pb
Nuquí	Sr	Ca, Ti,		Fe, Mg, Mn, Li	Cu, Zn, V		Co, Cr, Pb
Coqui	Sr	Ca	Ti	Fe, Mg, Mn, Li	V	Cu, Zn	Co, Cr, Pb

**Fig. 3.** I_{geo} values for sediments collected from different beaches along Bahía Solano and Nuquí, Department of Chocó, Colombia.

sediments of Bahía Solano and Nuquí are presented in Table 3. Values of < 1.5 for Sr suggest its crustal origin along with Ca and Ti. Fe, Mg, Mn and Li displayed EF values between 3 and 10, thus revealing a moderately severe enrichment of these metals. Severe and extremely severe enrichments of Co, Cr, Cu, Pb, Zn and V (EF > 50) reflect the extensive gold mining activities in Chocó (Yacoub et al., 2012; Wasiu et al., 2016).

As proposed by Muller (1979), another relative contamination assessment of sediments is the geoaccumulation index (I_{geo}), which is calculated by I_{geo} = (log₂ C_n/1.5 B_n), where C_n is the concentration of the element 'n', B_n is the background concentration for the element and

1.5 is a factor employed to compensate for the difference in metal concentrations due to lithogenic variations. On the basis of the respective ranges of I_{geo} values, seven classes (0–6) were obtained from uncontaminated to extremely contaminated sediments as shown in Fig. 3. Sediments were uncontaminated to moderately contaminated (Class 1) with Cu, Zn and V. In sediments from the El Valle beach, Cu alone presented moderate contamination (Class 2) and revealed the influence of substantial fishing activities (Nikolaou et al., 2014; Berillis et al., 2017). However, the heavy to extreme contamination of Co, Cr (Class 3) and Pb (Class 4) in the beach sediments was attributed to the extensive illegal mining activities in the Department of Chocó.

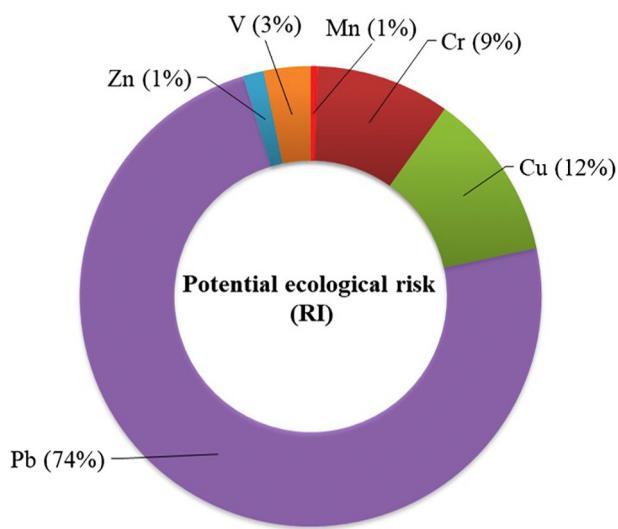


Fig. 4. Potential ecological risk index values calculated for all the sediments collected from the beaches along Bahía Solano and Nuquí, Department of Chocó, Colombia.

The PERI developed by Hakanson (1980) is one of the most widely applied methods to evaluate metal contamination and their associated ecological risk (Li et al., 2015; Maanan et al., 2015; Lin et al., 2016; Yang et al., 2016). It is calculated as follows: $RI = \sum_i^6 ER_f^i$; $ER_f^i = Tr^i \times C_f^i = Tr^i \times \left(\frac{C_s^i}{C_b^i} \right)$, where ER_f^i is the potential ecological risk factor for a given element i ; Tr^i is the biological toxicity factor for element i , which is defined as Mn = 1, Cr = 2, Cu = Pb = 5, Zn = 1 and V = 2 (Hakanson, 1980); and C_f^i , C_s^i and C_b^i are the contamination factor, the concentration in the sediment and the background reference value for element i , respectively. According to Hakanson (1980), four categories of PERI have been defined ranging from low ecological risk ($PERI \leq 150$) to significantly high ecological risk ($RI > 600$). From the PERI results obtained in the present study, the ranking of metals was in the following order: Pb (723) > Cu (116) > Cr (92) > V (32) > Zn (14) > Mn (5) (Fig. 4). All the metals in Bahía Solano and Nuquí sediments displayed a low ecological risk, except for Pb with a PERI value 723; this indicates the high ecological risk caused by Pb. In majority, Pb accounted for 74% (Fig. 4) of the ecological threat in beach sediments resulting from the atmospheric deposition of dust released during the crushing and grinding of the ore to extract gold (Tirima et al., 2017).

In summary, this study accounted data on the current state of metal concentrations and enrichments along the Pacific coast of Bahía Solano and Nuquí beach sediments. Briefly, the key findings include the presence of Co, Cr, Cu, Pb, Zn and V at elevated levels, among which Pb exhibits significant ecological risks. The main sources of these metals are illegal gold mining and extraction processes in the Department of Chocó. However, the lack of previous studies over this region limits the comprehension and estimation of metal enrichments that occurred over the last several decades. The presented results serve as a baseline for further investigations, and this study portrays the impacts of long-term mining practices on the marine ecosystem in Colombia.

Acknowledgement

HGM thanks the support of UTCH-Colombia (0011/2015) and COLCIENCIAS, 694-2014. MPJ wishes to express his thanks to IPN (COFAA, EDI), Mexico. MPJ and PDR thank Sistema Nacional de Investigadores (SNI), CONACyT, México. VCS and DMR thank CONACyT for the research fellowship. This article is part of the 'Blue Flag Beaches: Vision 2030', and it is the 101st contribution (partial) from Earth System Science Group (ESSG), Chennai, India. This article is also

part of the "Scientific Development Program" initiated by the University of Medellin, Colombia, in 2014.

References

- Ahmadi, A., Khezri, M., Abdollahzadeh, A.A., Askari, M., 2015. Bioleaching of copper, nickel and cobalt from the low grade sulfidic tailing of Golgohar Iron Mine, Iran. Hydrometallurgy 154, 1–8.
- Alexander, C.R., Smith, R.G., Calder, F.D., Schropp, S.J., Windom, H.L., 1993. The historical record of metal enrichment in two Florida estuaries. Estuarine 16 (36), 627–637.
- Alvarez, S., Jessick, A.M., Palacio, J.A., Kolok, A.S., 2012a. Methylmercury concentrations in six fish species from two Colombian rivers. Bull. Environ. Contam. Toxicol. 1–4.
- Alvarez, S., Kolok, A.S., Jimenez, L.F., Granados, C., Palacio, J.A., 2012b. Mercury concentrations in muscle and liver tissue of fish from marshes along the Magdalena River, Colombia. Bull. Environ. Contam. Toxicol. 1–5.
- Ashraf, A., Saion, E., Gharibshahi, E., Mohamed Kamari, H., Yap, C., Hamzah, M., Elias, M., 2016. Rare earth elements in core marine sediments of coastal East Malaysia by instrumental neutron activation analysis. Appl. Radiat. Isot. 107, 17–23.
- Avila, I.C., Garcia, C., Juan Carlos, B., 2008. A note on the use of dolphins as bait in the artisanal fisheries off Bahía Solano, Chocó, Colombia. J. Cetacean Res. Manag. 10 (2), 179–182.
- Berillies, P., Mente, E., Kormas, K.A., 2017. The use of copper alloy in aquaculture fish net pens: mechanical, economic and environmental advantages. J. Fish. Sci. 11 (4), 001–003.
- Bryan, C.G., Hallberg, K.B., Johnson, D.B., 2006. Mobilisation of metals in mineral tailings at the abandoned São Domingos copper mine (Portugal) by indigenous acidophilic bacteria. Hydrometallurgy 83, 184–194.
- Buat-Ménard, P., Chesselet, R., 1979. Variable influence of the atmospheric flux on the trace metal chemistry of oceanic suspended matter. Earth Planet. Sci. Lett. 42, 398–411.
- Chopra, A.K., Pathak, C., 2015. Accumulation of heavy metals in the vegetables grown in wastewater irrigated areas of Dehradun, India with reference to human health risk. Environ. Monit. Assess. 187, 445.
- Cordy, P., Veiga, M.M., Salih, I., Al-Saadi, S., Console, S., Garcia, O., Mesa, L.A., Velásquez-López, P.C., Roeser, M., 2011. Mercury contamination from artisanal gold mining in Antioquia, Colombia: the world's highest per capita mercury pollution. Sci. Total Environ. 410–411, 154–160.
- Correa, I.D., Morton, R.A., 2011. Coasts of Colombia. U.S. Department of the Interior USGS, St. Petersburg, Florida Available from. <http://coastal.er.usgs.gov/coastscolombia/>.
- El Nemr, A., El-Said, G.F., Khaled, A., Ragab, S., 2016. Distribution and ecological risk assessment of some heavy metals in coastal surface sediments along the Red Sea, Egypt. Int. J. Sed. Res. 31, 64–172.
- El-Sorogy, A.S., Tawfik, M., Almadani, S.A., Attiah, A., 2016. Assessment of toxic metals in coastal sediments of the Rosetta area, Mediterranean Sea, Egypt. Environ. Earth Sci. 75, 398.
- EPA Method 3051A, 2007. Microwave Assisted Acid Digestion of Sediments, Sludges, Solids and Oils. Revision 1, Feb, 2007, Washington, DC. pp. 1–30.
- Faggi, A., Dadon, J., 2011. Temporal and spatial changes in plant dune diversity in urban resorts. J. Coast. Conserv. 15, 585–594.
- Falagán, C., Grail, B.M., Johnson, D.B., 2017. New approaches for extracting and recovering metals from mine tailings. Miner. Eng. 106, 71–78.
- Fashola, M.O., Ngole-Jeme, V.M., Babalola, O.O., 2016. Heavy metal pollution from gold mines: environmental effects and bacterial strategies for resistance. Int. J. Environ. Res. Public Health 13 (11), 1047.
- Ferreira da Silva, E., Zhang, C., Serrano Pinto, L., Patinha, C., Reis, P., 2004. Hazard assessment on arsenic and lead in soils of Castromil gold mining area, Portugal. Appl. Geochem. 19, 887–898.
- Gao, X., Chen, C.T.A., 2012. Heavy metal pollution status in surface sediments of the coastal Bohai Bay. Water Res. 46 (6), 1901–1911.
- García, E.M., Cruz-Motta, J.J., Farina, O., Bastidas, C., 2008. Anthropogenic influences on heavy metals across marine habitats in the western coast of Venezuela. Cont. Shelf Res. 28, 2757–2766.
- Gómez Tapias, J., Almanza Meléndez, M.F., 2015. Mapa Geológico de Colombia: Servicio Geológico Colombiano. pp. 2694513.
- Goñi, E.A., Sabogal, A., Asmat, R., 2014. Minería informal aurífera en Colombia. Fedesarrollo, Bogotá. www.repository.fedesarrollo.org.co/bitstream/11445/368/3/Mineria%20informal%20aurifera%20en%20Colombia%20%20Informe_linea_base_mineria_informal%20-%20pagina%20Fedesarrollo.pdf.
- Gutiérrez-Mosquera, H., Sujitha, S.B., Jonathan, M.P., Sarkar, S.K., Medina-Mosquera, F., Ayala-Mosquera, H., Morales-Mira, G., Arreola-Mendoza, L., 2017. Mercury levels in human population from a mining district in Western Colombia. J. Environ. Sci. <http://dx.doi.org/10.1016/j.jes.2017.12.007>.
- Hakanson, L., 1980. An ecological risk assessment index for aquatic contamination control, a sedimentological approach. Water Res. 14, 975–1001.
- Ibrahim, S., Ashraf, A., Saion, E., Wood, A., Yusoff, W., 2014. Weathering Product of Granite as a Possible Source of Strategic Mineral, From Sources to Solution. Springer, pp. 497–502.
- Leyva, P., 1993. Colombia Pacífico. 2 Vols. Fondo Para la Protección del Medio Ambiente. José Celestino Mutis. Fondo FEN, Bogotá, DE.
- Li, C., Song, C., Yin, Y., Sun, M., Tao, P., Shao, M., 2015. Spatial distribution and risk assessment of heavy metals in sediments of Shuangtaizi Estuary, China. Mar. Pollut.

- Bull. 98, 358–364.
- Lin, Y., Meng, F., Du, Y., Tan, Y., 2016. Distribution, speciation, and ecological risk assessment of heavy metals in surface sediments of Jiaozhou Bay, China. *Hum. Ecol. Risk Assess.* 22, 1253–1267.
- Long, E.R., MacDonalnd, D.D., Smith, S., Calder, F., 1995. Incidence of adverse biological effects within ranges of chemical concentrations in marine and estuarine sediments. *Environ. Manag.* 19, 81–97.
- Maanan, M., Saddik, M., Maanan, M., Chaibi, M., Assobhei, O., Zourarah, B., 2015. Environmental and ecological risk assessment of heavy metals in sediments of Nador lagoon, Morocco. *Ecol. Indic.* 616–626.
- MacDonald, D.D., Ingersoll, C.G., Berger, T.A., 2000. Development and evaluation of consensus-based sediment quality guidelines for freshwater ecosystems. *Arch. Environ. Contam. Toxicol.* 39, 20–31.
- Maria Fernanda, Collazos-Santos, 2014. Definición de Objetivos de calidad de Vertimientos en la Bahía de Buenaventura desarrollo de la Fase I. Universidad Autónoma de Occidente, pp. 1–115 (M.Sc. Thesis).
- Marrugo-Negrete, J., Verbel, J.O., Ceballos, E.L., Benitez, L.N., 2008. Total mercury and methylmercury concentrations in fish from the Mojana region of Colombia. *Environ. Geochem. Health* 30 (1), 21–30.
- Matallana, M., 1999. Documento Proyecto Utría Regional PUR: Sociedad Economía y Territorio. Fundación Natura 34pp. Available from: www.natura.org.com.
- Muller, G., 1979. Heavy metals in the sediment of the Rhine—changes seity. 1971. *Umsch Wiss Tech.* 79, 778–783.
- Nagarajan, R., Jonathan, M.P., Roy, Priyadarsi D., Wai-Hwa, L., Prasanna, M.V., Sarkar, S.K., Navarrete-Lopez, M., 2013. Metal concentrations in sediments from tourist beaches of Miri City, Sarawak, Malaysia (Borneo Island). *Mar. Pollut. Bull.* 73, 369–373.
- Navarrete-López, M., Jonathan, M.P., Rodríguez-Espínosa, P.F., Salgado-Galeana, J.A., 2012. Autoclave decomposition method for metals in soils and sediments. *Environ. Monit. Assess.* 184, 2285–2293. Article online. <https://doi.org/10.1007/s10661-011-2117-4>.
- Nikolaou, M., Neofitou, N., Skordas, K., Castritsi-Catharios, J., Tziantziou, L., 2014. Fish farming and anti-fouling paints: a potential source of Cu and Zn in farmed fish. *Aquat. Environ. Interact.* 5, 163–171.
- Olivero, J., Johnson, B., Arguello, E., 2002. Human exposure to mercury in San Jorge river basin, Colombia (South America). *Sci. Total Environ.* 289 (1–3), 41–47.
- Palacios-Torres, Y., Caballero-Gallardo, K., Olivero-Verbel, J., 2018. Mercury pollution by gold mining in a global biodiversity hotspot, the Choco biogeographic region, Colombia. *Chemistry* 93, 421–430.
- Pappa, F.K., Tsabaris, C., Ioannidou, A., Patiris, D.L., Kaberi, H., Pashalidis, I., Eleftheriou, G., Androulakaki, E.G., Vlastou, R., 2016. Radioactivity and metal concentrations in marine sediments associated with mining activities in Ierissos Gulf, North Aegean Sea, Greece. *App. Radiat. Isot.* 116, 22–33.
- Parra, S., Bravo, M.A., Quiroz, W., Querol, X., Paipa, C., 2015. Distribution and pollution assessment of trace elements in marine sediments in the Quintero Bay (Chile). *Mar. Pollut. Bull.* 99 (1–2), 256–263.
- Pérez-Maqueo, O., Luisa Martínez, M., Nahuacatl, Rosendo Cóscatl, 2017. Is the protection of beach and dune vegetation compatible with tourism? *Tour. Manag.* 58, 175–183.
- PROEXPORT (Promoción de Turismo, Inversión y Exportaciones), 2011. Informe de rendición de cuentas para el sector comercio, industria y turismo. Ministerio de Comercio, Industria y Turismo, Bogota, Colombia.
- Propin-Frejómil, E., Sánchez Crispín, A., 2007. Tipología de los destinos turísticos preferenciales en México. *Cuad. Turismo* 19, 147–166.
- Ra, K., Kim, E.-S., Kim, K.-T., Kim, J.-K., Lee, J.-M., Choi, J.-Y., 2013. Assessment of heavy metal contamination and its ecological risk in the surface sediments along the coast of Korea. In: Conley, D.C., Masselink, G., Russell, P.E., O'Hare, T.J. (Eds.), Proceedings 12th International Coastal Symposium (Plymouth, England). *J Coast Res. Special Issue No.* 65pp. 105–110 (ISSN 0749-0208).
- Ramachandra, T.V., Sudarshan, P.B., Mahesh, M.K., Vinay, S., 2018. Spatial patterns of heavy metal accumulation in sediments and macrophytes of Bellandur wetland, Bangalore. *J. Environ. Manag.* 206, 1204–1210.
- Rangel-Buitrago, N., Williams, A., Anfuso, G., 2017. Hard protection structures as a principal coastal erosion management strategy along the Caribbean coast of Colombia. A chronicle of pitfalls. *Ocean Coast Manag.* 1–18.
- Retama, I., Jonathan, M.P., Roy, Priyadarsi D., Rodríguez-Espínosa, P.F., Nagarajan, R., Sarkar, S.K., Morales-García, S.S., Muñoz-Sevilla, N.P., 2016. Metal concentrations in sediments from tourist beaches of Huatulco, Oaxaca, Mexico: an evaluation of post-Easter week vacation. *Environ. Earth Sci.* 75, 375.
- Romano, E., Giudici, G.D., Bergamin, L., Andreucci, S., Maggi, C., Pierfranceschi, G., Magno, M.C., Ausili, A., 2017. The marine sedimentary record of natural and anthropogenic contribution from the Sulcis-Iglesiente mining district (Sardinia, Italy). *Mar. Pollut. Bull.* 122, 331–343.
- Rumisha, C., Elskens, M., Leermakers, M., Kochzius, M., 2012. Trace metal pollution and its influence on the community structure of soft bottom molluscs in intertidal areas of the Dar es Salaam coast, Tanzania. *Mar. Poll. Bull.* 64, 521–531.
- Salazar-Camacho, C., Salas-Moreno, M., Marrugo-Madrid, S., Marrugo-Negrete, J., Díez, S., 2017. Dietary human exposure to mercury in two artisanal small-scale gold mining communities of northwestern Colombia. *Environ. Int.* 107, 47–54.
- Santhiya, G., Lakshumanan, C., Jonathan, M.P., Roy, P.D., Navarrete-Lopez, M., Srinivasulu, S., Uma-Maheswari, B., Krishnakumar, P., 2011. Metal enrichment in beach sediments from Chennai Metropolis, SE coast of India. *Mar. Pollut. Bull.* 62 (11), 2537–2542.
- Schropp, S.J., Lewis, F.G., Windom, H.L., Ryan, J.D., Calder, F.D., Burney, L.C., 1990. Interpretation of metal concentrations in estuarine sediments of Florida using aluminum as a reference element. *Estuarine* 13 (3), 227–235.
- Tirima, S., Bartrem, C., Lindern, I., Braun, M., Lind, D., Anka, S.M., Abdullahi, A., 2017. Food contamination as a pathway for lead exposure in children during the 2010–2013 lead poisoning epidemic in Zamfara, Nigeria. *J. Environ. Sci.* <http://dx.doi.org/10.1016/j.jes.2017.09.007>.
- United States Environmental Protection Agency (USPEA), 2001. The Role of Screening Level Risk Assessments and Refining Contaminants of Concern in Baseline Ecological Risk Assessments Publications. 9345-014, EPA 540/F-01/14. (June 2001).
- Vallejo Toro, P.P., Vásquez Bedoya, L.F., Darío Correa, I., Bernal Franco, G.R., Alcántara-Carrió, J., Palacio Baena, J.A., 2016. Impact of terrestrial mining and intensive agriculture in pollution of estuarine surface sediments: spatial distribution of trace metals in the Gulf of Urabá, Colombia. *Mar. Pollut. Bull.* 111, 311–320.
- Vetrimurugana, E., Shruti, V.C., Jonathan, M.P., Roy, Priyadarsi D., Rawlins, B.K., Rivera-Rivera, D.M., 2018. Metals and their ecological impact on beach sediments near the marine protected sites of Sodwana Bay and St. Lucia, South Africa. *Mar. Pollut. Bull.* 127, 568–575.
- Vidinha, J.M., Rocha, F., Silva, E., Patinha, C., Andrade, C., 2009. Geochemical beach sediments studies—a contribution to a standard definition useful for public health. *J. Coast. Res.* 56, 905–908.
- Vikas, M., Dwarkish, G.S., 2015. Coastal pollution: a review. *Aquat. Proc.* 4, 381–388.
- Vink, R., Behrendt, H., Salomons, W., 1999. Development of the heavy metal pollution trends in several European rivers: an analysis of point and diffuse sources. *Water Sci. Technol.* 39 (12), 215–223.
- Wang, L., Qiang, Li, Hongsheng, Bi, Xian-zhong, M., 2016. Human impacts and changes in the coastal waters of south China. *Sci. Total Environ.* 562, 108–114.
- Wasiu, M.O., Ayodele, O.E., Ayodele, T.I., Oluremi, O.I., Temitope, O.K., Temitope, F.O., 2016. Heavy metal contamination in stream water and sediments of gold mining areas of South Western Nigeria. *Afr. J. Environ. Sci. Tech.* 10 (5), 150–161.
- Wedepohl, H., 1995. The composition of the continental crust. *Geochim. Cosmochim. Acta* 59, 1217–1239.
- Windom, H.L., Schropp, S.J., Calder, F.D., Ryan, J.D., Smith, R.G., Burney, L.C., Lewis, F.G., Rawlinson, C.H., 1989. Natural trace-metal concentrations in estuarine and coastal marine-sediments of the southeastern United-States. *Environ. Sci. Technol.* 23 (3), 314–320.
- Xu, X., Cao, Z., Zhang, Z., Li, R., Hu, B., 2016. Spatial distribution and pollution assessment of heavy metals in the surface sediments of the Bohai and Yellow seas. *Mar. Pollut. Bull.* 110, 596–602.
- Yacoub, C., Pérez-Foguet, A., Miralles, N., 2012. Trace metal content of sediments close to mine sites in the Andean region. *Sci. World J.* 2012, 732519 12 pages. <https://doi.org/10.1100/2012/732519>.
- Yang, Q., Hu, G., Yu, R., He, H., Lin, C., 2016. Distribution, fractionation, and contamination assessment of heavy metals in offshore surface sediments from western Xiamen Bay, China. *Acta Geochim.* 35, 355–367.
- Ye, X., Xiao, W., Zhang, Y., Zhao, S., Wang, G., Zhang, Q., Wang, Q., 2015. Assessment of heavy metal pollution in vegetables and relationships with soil heavy metal distribution in Zhejiang province, China. *Environ. Monit. Assess.* 187, 378.
- Youssef, M., El-Sorogy, A.S., Al-Kahtany, K.H., Al-Otaibi, N., 2015. Environmental assessment of coastal surface sediments Tarut Island, Arabian Gulf (Saudi Arabia). *Mar. Poll. Bull.* 96, 424–433.